## Accepted Manuscript

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PII: S0269-7491(18)34771-7

DOI: https://doi.org/10.1016/j.envpol.2019.05.117

Reference: ENPO 12662

To appear in: Environmental Pollution

Received Date: 24 October 2018

Revised Date: 22 May 2019

Accepted Date: 22 May 2019

Please cite this article as: Sheng Ng, C.F., Hashizume, M., Obase, Y., Doi, M., Tamura, K., Tomari, S., Kawano, T., Fukushima, C., Matsuse, H., Chung, Y., Kim, Y., Kunimitsu, K., Kohno, S., Mukae, H., Associations of chemical composition and sources of PM<sub>2.5</sub> with lung function of severe asthmatic adults in a low air pollution environment of urban Nagasaki, Japan, *Environmental Pollution* (2019), doi: https://doi.org/10.1016/j.envpol.2019.05.117.

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PM<sub>2.5</sub> Constituent Concentration (ng/m<sup>3</sup>)





X Associated with lower PEF (from personal monthly max), and robust in 2-pollutant model
 X Higher odds of lung function decline (≥15% within 1-week), but not robust in 2-pollutant model

# Associations of chemical composition and sources of $PM_{2.5}$ with lung function of severe asthmatic adults in a low air pollution environment of urban Nagasaki, Japan

Chris Fook Sheng Ng<sup>a,b,c</sup>, Masahiro Hashizume<sup>a,c</sup>, Yasushi Obase<sup>d,\*</sup>, Masataka Doi<sup>e</sup>, Kei Tamura<sup>f</sup>, Shinya Tomari<sup>g</sup>, Tetsuya Kawano<sup>d</sup>, Chizu Fukushima<sup>d</sup>, Hiroto Matsuse<sup>h</sup>, Yeonseung Chung<sup>i</sup>, Yoonhee Kim<sup>j</sup>, Kenichi Kunimitsu<sup>e</sup>, Shigeru Kohno<sup>d</sup>, Hiroshi Mukae<sup>d</sup>

<sup>a</sup> School of Tropical Medicine and Global Health, Nagasaki University, Nagasaki, Japan

<sup>b</sup> Department of Public Health, Environment and Society, London School of Hygiene &

Tropical Medicine, London, United Kingdom

<sup>c</sup> Institute of Tropical Medicine, Nagasaki University, Nagasaki, Japan

<sup>d</sup> Department of Respiratory Medicine, Nagasaki University Graduate School of Biomedical Sciences, Nagasaki, Japan

<sup>e</sup> Nagasaki Prefectural Institute of Environment and Public Health, Omura, Nagasaki, Japan

<sup>f</sup> Environmental Policy Division, Nagasaki Prefectural Government, Nagasaki, Japan

<sup>g</sup> Department of Respiratory Medicine, Isahaya General Hospital, Japan Community Health care Organization (JCHO), Isahaya, Nagasaki, Japan

<sup>h</sup> Department of Internal Medicine, Division of Respiratory Medicine, Toho University Ohashi Medical Center, Tokyo, Japan

<sup>i</sup> Department of Mathematical Sciences, Korea Advanced Institute of Science and Technology, Daejeon, South Korea

<sup>j</sup> Department of Global Environmental Health, Graduate School of Medicine, The University of Tokyo, Tokyo, Japan

\* Corresponding author at:
Department of Respiratory Medicine, Graduate School of Biomedical Sciences, Nagasaki
University, 1-7-1 Sakamoto, Nagasaki 852-8501, Nagasaki, Japan
E-mail address: obaseya@nagasaki-u.ac.jp

#### Abstract

Previous studies have linked ambient PM2.5 to decreased pulmonary function, but the influence of specific chemical elements and emission sources on the severe asthmatic is not well understood. We examined the mass, chemical constituents, and sources of PM<sub>2.5</sub> for short-term associations with the pulmonary function of adults with severe asthma in a low air pollution environment in urban Nagasaki, Japan. We recruited 35 asthmatic adults and obtained the daily record of morning peak expiratory flow (PEF) in spring 2014-2016. PM<sub>2.5</sub> filters were extracted from an air quality monitoring station (178 days) and measured for 27 chemical elements. Source apportionment was performed using Positive Matrix Factorization (PMF). We fitted generalized linear model with generalized estimating equation (GEE) method to estimate changes in PEF (from personal monthly maximum) and odds of severe respiratory deterioration (first  $\geq$  15% PEF reduction within a 1-week interval) associated with mass, constituents, and sources of PM<sub>2.5</sub>, with adjustment for temperature and relative humidity. Constituent sulfate  $(SO_4^{2-})$  and  $PM_{2.5}$  from oil combustion and traffic were associated with reduced PEF. An interquartile range (IQR) increase in  $SO_4^{2-}$  (3.7 µg/m<sup>3</sup>, average lags 0-1) was associated with a decrease of 0.38% (95% confidence interval = -0.75% to -0.001%). An IQR increase in oil combustion and traffic-sourced PM<sub>2.5</sub> (2.64  $\mu$ g/m<sup>3</sup>, lag 1) was associated with a decrease of 0.33% (-0.62% to -0.002%). We found a larger PEF decrease associated with PM<sub>2.5</sub> from dust/soil on Asian Dust days. There was no evidence linking total mass and metals to reduced pulmonary function. The ventilatory capacity of adults with severe asthma is susceptible to specific constituents/sources of PM<sub>2.5</sub> such as sulfate and oil combustion and traffic despite active self-management of asthma and low air pollution levels in the study location.

Keywords: particulate matter; sulfate; oil combustion & traffic; dust; spirometry; asthma

*Capsule:* Sulfate aerosols and  $PM_{2.5}$  from oil combustion and traffic negatively affect lung function of adults with severe asthma in low  $PM_{2.5}$  area.

#### 1 **1.** Introduction

2

3 Inverse association between ambient particulate matter with aerodynamic diameter of  $\leq 2.5$ 4 µm (PM<sub>2.5</sub>) and respiratory morbidity has been widely documented (Xu et al., 2016; Bell et 5 al., 2009; Bell et al., 2014). Several studies have also documented a negative association between PM<sub>2.5</sub> and lung function, but the reported risk estimates varied geographically and 6 7 seasonally (Lagorio et al., 2006; Ma et al., 2008; Wiwatanadate and Liwsrisakun, 2011; Strak 8 et al., 2012; Wu et al., 2014a; Yamazaki et al., 2011). It has been suggested that this 9 heterogeneity could be explained by the mixture of PM<sub>2.5</sub> chemical constituents, which varied 10 depending on the type of emission sources and their temporal characteristic, thereby resulting 11 in different toxicity across study locations and season (Bell et al., 2007; Hong et al., 2007). 12 These observations highlight the importance to understand how specific chemical 13 constituents and emission sources of PM<sub>2.5</sub> affect the lung functions of populations in 14 different locations and time period. But research in this area remains limited, especially in 15 low concentration environments such as Japan.

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17 There is also a lack of understanding of how lung function of severe asthmatic individuals, a 18 highly susceptible subgroup, can be influenced by certain chemical species or sources of 19 PM<sub>2.5</sub>. Evidence of association between the overall mass of PM<sub>2.5</sub>/PM<sub>10</sub> and lung function has 20 been indicated previously in healthy individuals (Chang et al., 2012; Zwozdziak et al., 2016) 21 and the asthmatic (Lagorio et al., 2006; Liu et al., 2009; Ma et al., 2008; Odajima et al., 2008; 22 Wiwatanadate and Liwsrisakun 2011; Yamazaki et al., 2011). Specific chemical constituents 23 of PM<sub>2.5</sub> and their emission sources have been linked to reduced pulmonary functions in 24 healthy adults (Wu et al., 2013; Cakmak et al., 2014; Wu et al., 2014b). In addition, a study 25 has demonstrated that children without the use of inhaled corticosteroids recorded larger risk 26 estimates associated with exposure to PM<sub>2.5</sub> total mass, implying potential effect modification 27 by treatment (Liu et al., 2009). Research is therefore pertinent to provide a comprehensive 28 chemical and source-specific characterization of  $PM_{2.5}$  exposure among the severe asthmatic.

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The current study examines the associations of  $PM_{2.5}$  total mass, its chemical constituents, and sources with the lung function of adults diagnosed with severe asthma with ongoing selfmanagement in an urban area in Nagasaki prefecture, Japan. The study focuses on the spring season in consideration of the susceptibility of the region to transported air pollutants including Asian Dust (AD) facilitated by the predominantly westerly winds from the Asia

continent during this period (Guo et al., 2017). Furthermore, the relatively low concentration of  $PM_{2.5}$  (daily mean [SD] during the study period = 18.5 [8.3] µg/m<sup>3</sup>; daily mean of 35 µg/m<sup>3</sup> is the air quality standard in Japan) provides a unique opportunity to study the sensitivity of severe asthmatic adults exposed to a low ambient level.

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#### 40 **2.** Materials and Methods

#### 41 2.1. Study participants and health measurements

A panel study was conducted to investigate the short-term association of the chemical 42 43 constituents and sources of PM<sub>2.5</sub> with the lung function of adults clinically diagnosed with 44 severe asthma. We recruited patients from Nagasaki University Hospital and Isahaya General 45 Hospital (Fig. S1). Patients diagnosed with chronic obstructive pulmonary disease or enrolled 46 in clinical trials at the time of recruitment were deemed ineligible and excluded. We 47 administered questionnaires to elicit background information such as age, sex, smoking status, disease condition, and existing cardiovascular or respiratory comorbidities. Lung 48 49 function data were collected prospectively between the year 2014 and 2016, during which participants were instructed to measure their peak expiratory flow rate (PEF, L/min) using a 50 portable peak flow meter (Mini-Wright<sup>TM</sup>, Clement Clarke, Edinburgh, UK) twice a day – 51 morning and evening – and record the data in a personal diary. We extracted the lung 52 function data for spring season (March to May), a period coinciding with occasional AD 53 54 events that affect air quality in the region. For data analysis, the following exclusion criteria 55 were applied: participant with home address outside Nagasaki city or Isahaya city (n=1), 56 smokers (n=5), elderly subjects over 65 years of age (n = 25), and subjects with excessive 57 missing data (n = 8), leaving a total of 35 adults (30 from Nagasaki city and 5 from Isahaya 58 city) for the final analysis.

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We collected written informed consent from all participants before the study, the protocol for
which was approved by the clinical research ethics committee of Nagasaki University
Hospital (No. 13062413).

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#### 64 2.2. Environmental data

We obtained daily data on  $PM_{2.5}$  and its chemical constituents from the Nagasaki Prefectural Institute of Environment and Public Health. Measurements of outdoor  $PM_{2.5}$  ( $\mu$ g/m<sup>3</sup>) were provided by two continuous monitoring stations located near the hospitals in Nagasaki city

68 and Isahaya city (Fig. S1). Readings from the two stations were highly correlated (r = 0.96). To obtain the chemical elements of  $PM_{2.5}$ , we extracted the polytetrafluoroethylene (PTFE) 69 70 and quartz filters from the monitoring station at Isahaya city (Whatman PM<sub>2.5</sub> PTFE 71 membrane filters and Pallflex Air Monitoring 2500 QAT-UP filters). The filters were 72 automatically replaced at noon every day and collected once a week for analysis. To prevent 73 contamination and improve quantification, field blank filters subject to the same condition of 74 sample collection were deployed, and blank concentrations were used to revise all daily 75 measurements. Filter samples were available on 178 days (March 1 to March 30 in 2014, 76 March 2 to May 31 in 2015, and March 1 to April 26 in 2016 for a total of approximately 6 77 months). The carbonaceous fractions in the dust particles, organic carbon (OC) and elemental carbon (EC), were measured using the thermal optical reflectance method (Lab OC-EC 78 79 Aerosol Analyzer, Sunset Laboratory Inc., OR, USA). The ionic components of PM<sub>2.5</sub> such as sulfate (SO<sub>4</sub><sup>2-</sup>), nitrate (NO<sub>3</sub><sup>-</sup>), ammonium (NH<sub>4</sub><sup>+</sup>), sodium ion (Na<sup>+</sup>), potassium ion (K<sup>+</sup>), 80 magnesium ion  $(Mg^{2+})$ , and calcium ion  $(Ca^{2+})$  were determined using ion chromatography; 81 82 metal components such as aluminium (Al), arsenic (As), barium (Ba), cadmium (Cd), cerium 83 (Ce), cobalt (Co), copper (Cu), iron (Fe), lanthanum (La), manganese (Mn), molybdenum 84 (Mo), nickel (Ni), lead (Pb), rubidium (Rb), selenium (Se), titanium (Ti), vanadium (V), zinc 85 (Zn) were determined using inductively coupled plasma mass spectrometry (ICP-MS). For concentrations below the limit of detection, we substituted the data with values equivalent to 86 87 half the corresponding detection limit (DL). Chemical species with over 80% of values below DL were excluded from further analysis. 88

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#### 94 2.3. Statistical Analyses

We examined PEF measured in the morning because  $PM_{2.5}$  chemical constituents were measured from noon to noon. To account for the heterogeneity of lung function across participants, daily individual PEF readings were converted into percent deviations from the personal monthly maximum PEF given as  $[(x_t - \max\{x_{m1},..., x_{mn}\})*100/x_t]$ , where  $x_t$ represents the PEF value on day *t* in a given year, *m* represents the corresponding month for day *t*, and *n* is the number of days in month *m*. In addition, to indicate a sign of possible

We obtained daily 24-h averages of air temperature (°C) and relative humidity (%) from the Japan Meteorological Agency (JMA). Temperatures were measured at two stations (r = 0.99), while relative humidity at one station in Nagasaki city (Fig. S1).

101 asthma, we created a binary outcome variable to represent significant respiratory 102 deterioration, computed as 15% or larger reduction in PEF from the personal monthly 103 maximum. The cut-off value was recommended based on expert opinion with consideration 104 to the ongoing self-management of asthma by patients to prevent development of an 105 exacerbation during the study. To exclude repeated deteriorations that might be correlated, 106 we considered only the first incidence (morning or evening) within a rolling 1-week period.

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Positive matrix factorization (PMF) receptor model was applied to determine the major 108 109 sources of PM<sub>2.5</sub> and their daily contributions to chemical constituents at the study location. Analyses were performed using the US EPA PMF 5.0 (United States Environmental 110 111 Protection Agency). A total of 27 chemical species were included in the PMF analysis. We 112 down-weighted chemical species with weak correlation between the observed and predicted 113 values (Table S1). For each species, measurements below the DL were replaced with values 114 half the corresponding DL and given an uncertainty of DL\*5/6. Missing measurements were replaced with year-specific median and given an uncertainty four times the replacement 115 value. For all other measurements, uncertainty was computed as  $0.05^*x_i + DL_i$  on sample *i* of 116 117 a particular species for n=178 samples collected (Ito et al., 2004). We selected 6 to 9 factors 118 initially and produced multiple PMF solutions for comparison with existing results from 119 related studies (Suzuki et al., 2014; personal communication with Nagasaki Prefectural 120 Institute of Environment and Public Health). A total of 6 sources were identified after multiple base and bootstrap runs. To minimize rotational ambiguity, we performed factor 121 122 rotation and the final model has an FPEAK value of -0.1 with a small increase (0.07%) in the 123 Q-value (robust). To evaluate the variability of PMF solutions, we performed 500 bootstrap 124 runs and computed interval ratio for each species in a factor (Brown et al, 2015). The final 125 model was then used to estimate the daily levels of PM<sub>2.5</sub> sources and chemical constituents. 126

127 We fitted generalized linear model (GLM) with generalized estimating equation (GEE) method to handle the within-subject correlation using a first-order autoregressive correlation 128 structure (AR1) to investigate the short-term associations of PM<sub>2.5</sub>, its chemical constituents 129 130 and sources with the decline of PEF (using an identity link function), or the events of severe 131 respiratory deterioration (using a logit link function). Chemical constituents that were potentially harmful based on previous studies were included (Bell et al., 2014; Hong et al., 132 2010; Lagorio et al., 2006; Wu et al., 2014b) – OC, EC, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>, NO<sub>3</sub><sup>-</sup>, Al, As, Cd, Cu, 133 134 Fe, K, Mn, Ni, Pb, Ti, V, and Zn. For each chemical constituent and source, we examined the

135 single- and multi-day average exposure from lag 0 to 2. Temporal trend in the time series (nearly 6 months) was accounted for using a natural cubic spline of calendar time with 7 136 137 degrees of freedom. We included the day of the week, daily mean temperature and relative 138 humidity (both at lag 1), age and sex for adjustment. For constituents or sources that showed 139 evidence of associations in a single-pollutant model (*p*-value < 0.10), we checked their 140 sensitivity to the other particles or sources by including an adjustment term representing the 141 concentration of total mass. We also checked the sensitivity of the estimates to existing 142 cardiovascular and respiratory diseases. To examine possible effect modification by AD, we extracted the information of AD events reported by JMA (visibility-based) and using an 143 interaction term, we estimated the changes of PEF associated with PM2.5 total mass and PMF-144 145 resolved sources on the AD and Non-AD days.

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147 Results were reported as percentage changes of PEF from the personal monthly maximum or 148 as odds ratios (OR) of the first severe respiratory deterioration within a rolling 1-week 149 interval and the corresponding 95% confidence intervals for an interquartile (IQR) increase in 150 the level of  $PM_{2.5}$  chemical constituents or sources (Table 2). The GLM modeling with GEE 151 method was performed in R version 3.3.3 (R Core Team, 2017) using the *geepack* package 152 (Højsgaard et al., 2006).

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#### 154 **3. Results**

Thirty five asthmatic adults aged 20 to 65 years were analyzed, providing an average of 208 days of lung function measurements. The average PEF for males and females was 499 L/min and 329 L/min, respectively (Table 1). The largest PEF decrease from the personal monthly maximum was approximately 46% and 52% for males and females, respectively. The frequency of severe respiratory deterioration within a 1-week rolling period ranged from 0-10 times at an average of 3 events per person.

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Table 2 summarizes the daily level of  $PM_{2.5}$  total mass, composition by carbon, ion, metal, contribution by sources and weather variables. Daily  $PM_{2.5}$  averaged 18.5 µg/m<sup>3</sup> and was composed of mainly  $SO_4^{2-}$  (29.4%), OC (15.2%),  $NH_4^+$  (12.8%), EC (6.3%), and  $NO_3^-$ (5.1%). PMF analysis resolved 6 profiles to describe the possible sources of  $PM_{2.5}$  (Fig. S2). Bootstrap resamples reproduced 100% of all base profiles, except the oil combustion and traffic profile at 89.6%. Interval ratios for the key species in each source profile were generally low, except for those in the dust and soil profile (Fig. S3). The largest contribution

was from secondary sulfate (38.3%), followed by dust and soil (18.5%), oil combustion and
traffic (12.4%), and secondary nitrate (11.9%). The concentrations and proportions of
chemical species by source profile are described in Table S1 and Fig. S2.

172

173 Fig. 1 shows the estimated effects of PM<sub>2.5</sub> total mass and chemical constituents on PEF or the occurrence of severe respiratory deterioration. There was weak evidence linking  $SO_4^{2-}$ 174 and  $NH_4^+$  to a decrease of PEF at lag 1. When averaged across lags 0-1, an increase in  $SO_4^{2-}$ 175 was associated with 0.38% reduction in PEF (95% CI: -0.75%, -0.001% for an IQR increase 176 of 3.7  $\mu$ g/m<sup>3</sup>, Table S2). We observed weak evidence suggesting increased odds of severe 177 respiratory decline for increases in  $PM_{2.5}$  and  $NO_3^-$  at lag 0. For OC, an increase in the 2-day 178 179 average level was associated with higher odds of severe respiratory decline (OR: 1.58; 95% CI: 1.002, 2.51 for an IQR increase of 1.8  $\mu$ g/m<sup>3</sup>, Table S3). 180

181

182 Fig. 2 shows the estimated effects of PMF-resolved sources of PM<sub>2.5</sub> on lung functions. An increase in oil combustion and traffic source-specific PM<sub>2.5</sub> level was associated with 0.33% 183 reduction in PEF at lag 1 (95% CI: -0.62%, -0.002% for an IQR increase of 2.64 µg/m<sup>3</sup>, 184 Table S2). A weak negative association with PEF was observed for the secondary sulfate 185 source-apportioned PM<sub>2.5</sub> averaged across lags 0-1. This source was also associated with 186 higher odds of severe respiratory deterioration at lag 1 (OR: 1.59; 95% CI: 1.18, 2.13 for an 187 IQR increase of 6.65  $\mu$ g/m<sup>3</sup>), with larger odds estimated for multi-day exposure, i.e. average 188 lag 0-1 (Table S3). 189

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The results of multi-pollutant models are shown in Tables S4 and S5. The significant reductions in PEF associated with sulfate component and the oil combustion and traffic source were fairly insensitive to adjustment for other constituents (Table S4). The significant odds of severe respiratory deterioration associated with OC and the secondary sulfate source were sensitive to adjustment (Table S5). Additional adjustment for existing cardiovascular and/or respiratory diseases had negligible effect on the risk estimates (Table S6 and S7).

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Table S8 summarizes the total mass and PMF-resolved sources of  $PM_{2.5}$  by the occurrence of AD. Except for the secondary sulfate source, all other air pollutants showed higher concentrations on AD days. There was evidence of effect modification by AD; the estimated effect of  $PM_{2.5}$  from dust and soil was larger on AD days (Table S9).

#### 203 **4. Discussion**

This study examined the associations of ambient PM<sub>2.5</sub>, its chemical constituents and sources 204 205 with the lung function of severe asthmatic adults during the spring season in Nagasaki and Isahaya city located on Kyushu island in southern Japan. We found that  $SO_4^{2-}$  and  $PM_{2.5}$  from 206 oil combustion & traffic source were associated with reduced PEF in adults diagnosed with 207 208 severe asthma. The associations were observed on the preceding day (lag 1) and remained 209 significant in multi-pollutant models. We also found increased odds of severe respiratory 210 deterioration attributable to increases in OC and PM<sub>2.5</sub> apportioned to secondary sulfate source, but these associations were attenuated in a multi-pollutant setting. There was weak 211 evidence linking PM<sub>2.5</sub> total mass to the higher odds of severe respiratory deterioration. None 212 213 of the metals selected for analysis significantly influenced the lung function of study subjects. We also did not find evidence of effect modification by AD events during the study period. 214

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Results of our study suggest that certain chemical constituents of PM<sub>2.5</sub> are likely more 216 217 harmful than others in adults with severe asthma. In the current study, the ambient level of  $PM_{2.5}$  was low (mean [SD] of 18.5 [8.3]  $\mu$ g/m<sup>3</sup>), well within the air quality standard in Japan 218 219 (daily mean of 35  $\mu$ g/m<sup>3</sup> or less). We did not find significant evidence of reduced lung function attributable to  $PM_{2.5}$  total mass. But the two dominant species,  $SO_4^{2-}$  and OC, 220 221 demonstrated inverse associations. We estimated a very small change in PEF associated with  $SO_4^{2-}$ , and the association was robust to the adjustment for other species. In contrast, the 222 estimated effect of OC was sensitive to the same adjustment likely because of collinearity 223 224 with species originating from the same source (Fig. S2). A panel study of 21 healthy college 225 students in Beijing, China, reported comparable findings, but at higher exposure levels (medians [IQR] of PM<sub>2.5</sub> mass and sulfate were 57.3 [63.4]  $\mu$ g/m<sup>3</sup> and 6.6 [15.3]  $\mu$ g/m<sup>3</sup>, 226 227 respectively) (Wu et al., 2013). The study found an inverse association between sulfate and PEF with an estimated reduction of 0.99% (95% CI: -1.67%, -0.32%) in evening PEF per 228 IQR increase in the 3-day average concentration of the pollutant (15.3  $\mu$ g/m<sup>3</sup>). The study also 229 230 reported an association of  $PM_{2.5}$  total mass with  $FEV_1$ , but not with PEF. The inconsistency 231 has been noted elsewhere (Strak et al., 2012), and might be related to the lower sensitivity of 232 PEF in detecting a small change in the ventilatory function (Giannini et al., 1997). It has also 233 been suggested that PEF measurement requires more effort from subjects (Thiadens et al., 234 1999), and is more error-prone and less reliable (Wu et al., 2013). The underlying mechanism linking sulfate to lung function at low concentration is not well understood. Exposure may 235 236 impact pulmonary mechanical function including airway responsiveness, but the reported

effects on asthmatics so far have been inconsistent especially at low air pollutant concentrations (Schlesinger and Cassee, 2003). Sulfate has been implicated to influence the bioavailability of certain metallic species, which contribute to the cytotoxicity of PM (Reiss et al., 2007), but further investigation is necessary to address the association at lower exposure level.

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243 Secondary sulfate emission source was associated with higher odds of severe lung function 244 deterioration in single-pollutant model. The source was marked by high loadings of  $SO_4^{2-}$ ,  $NH_4^+$ , OC, and EC (Table S1 and Fig. S2) and was the largest contributor to ambient  $PM_{2.5}$ 245 (38.3%) at the study location. The estimated effect for this source was likely driven by OC 246 247 for which an association was observed independently (Table S3). Relatively similar source 248 profile has been documented in Seoul (Heo et al., 2009) and Taiwan (Gugamsetty et al., 249 2012). Studies that reported this source described it as common near industrialized areas where there were high concentrations of sulfur dioxide (SO<sub>2</sub>), an important precursor for the 250 251 formation of secondary sulfates through photochemical or chemical processes (Heo et al., 2009; Gugamsetty et al., 2012; Oguilei et al., 2006; Wu et al., 2014b). In Beijing for 252 example, coal-burning industrial facilities have been implicated as a major source of SO<sub>2</sub> 253 254 contributing to sulfate level in the city (Wu et al., 2014b). In Baltimore, Maryland, USA, 255 coal-fired power plants have been linked to high sulfate loading through the analysis of wind 256 direction, which demonstrates the possibility of long-range transport (Oguilei et al., 2006). 257 The mixing of local and transported secondary sulfate was also observed in a study in Detroit, Michigan, USA (Morishita et al., 2006), where high sulfur concentration exceeding 258 259 the capacity of local sources was recorded, indicating the possibility of transported secondary 260 sulfate particles from distant industrial sources and power plants. In Seoul, Korea, a multi-261 year source-apportionment study has suggested two possible sources of transported sulfate 262 aerosols that affected the city – the industrial areas on eastern coastal China and emissions 263 from the ships on Yellow Sea (Heo at al., 2009). Indeed, findings from a study conducted at 264 Fukuoka city and Fukue island in Japan supported the supposition of long-range transport (Suzuki et al., 2014). The study showed that aerosol concentrations at the two study sites 265 266 were affected by coal combustion emissions originating from continental Asia, as well as by emissions from marine ships. These findings imply susceptibility of the study location to both 267 268 local and transported sulfate aerosols in the region given the proximity and the downwind travel of continental air masses. 269

271 Oil combustion and traffic source showed a consistent association with reduced ventilatory capacity in adults with severe asthma. This source was characterized by high loadings of V, 272 273 Ni, Se, and Mo (Fig. S2). Previous studies conducted in Beijing (Yu et al., 2013) and Taiwan 274 (Gugamsetty et al., 2012) have reported comparable source profile with the exception of Mo. 275 V and Ni are known tracers for oil combustion and traffic emissions, respectively. Some 276 known examples of oil combustion sources are oil-fired power plants, ships, and heating oil 277 (Lee et al., 2011; Bove et al., 2016), while for traffic emission, gasoline and diesel engines 278 are commonly reported (Lin et al., 2015; Shafer et al., 2011). Mo was not commonly 279 observed in this source profile but when highly correlated with Cu, might indicate wear 280 debris from road traffic as suggested in a study of traffic-derived particles (Lin et al., 2015). 281 In the current study, the correlation between the two was high (Spearman's  $\rho = 0.85$ ) 282 implying road traffic related contributions. Se has been attributed to coal combustion 283 (Morishita et al., 2006), but in the current study a large proportion of the element was also 284 found in another PMF-resolved source profile labeled as industrial emission source (Fig. S2), 285 which did not exhibit a relationship with lung function.

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Whereas a negative association was observed with the oil combustion and traffic source, we 287 288 did not detect similar association with the individual trace elements, i.e. V and Ni. A possible 289 explanation is the variety of sources producing these elements. For example, it was noted 290 previously that besides traffic/oil combustion, V and Ni could also be emitted by marine 291 vessels and industrial sources (de Foy et al., 2012). Consequently, PMF-resolved source-292 specific particles might show an association when the individual element aggregated from 293 multiple sources does not, or vice-versa (Bell et al. 2014). Other possible explanations 294 include inconsistency in measurement quality and uncertainties in the source-apportionment 295 method and the measures of constituents. Notwithstanding, V and Ni have been previously 296 linked to negative respiratory health outcomes. A study based on a birth cohort recruited in 297 New York city between 1998 and 2006 found that V and Ni were associated with wheezing 298 symptoms among children aged 2 years or younger (Patel et al., 2009). A multi-city study 299 using data from 106 counties in the US reported an increased risk of respiratory 300 hospitalization associated with the same-day concentration of V and Ni (Bell et al., 2009). 301 Another US study based on Medicare data comprising populations aged 65 years and over 302 observed similar short-term associations for both chemical species (Bell et al., 2014). On the 303 contrary, a study in Hong Kong did not find any association between the elements and 304 respiratory emergency hospitalizations (Pun et al., 2014).

306 We found evidence suggesting a larger lung function decline was associated with dust and 307 soil source-related PM<sub>2.5</sub> on AD days (Table S9). This was not observed with PM<sub>2.5</sub> mass or 308 other sources. A study conducted at a nearby city of Fukuoka to study the relationship 309 between suspended particulate matter (SPM) and asthma hospitalization of children under 12 310 years old did not find evidence of effect modification by AD (Ueda et al., 2010). In Seoul, 311 Hong et al. (2010) found that AD did not substantially modify the potency of metals on the 312 lung function of schoolchildren. Different definitions of AD have been used. For example, 313 Nakamura et al. (2015) utilized multiple criteria such as the concentration of SPM, Light 314 Detection and Ranging (LIDAR) extinction coefficient, and the correlation between SPM and 315 LIDAR. The authors reported an association between AD days and emergency department 316 visits for bronchial asthma and other respiratory diseases among children at a medical facility 317 in Nagasaki. Kanatani et al. (2010) conducted a study in western Japan using LIDAR 318 monitoring of mineral dust particles (non-spherical particles, a marker of AD events) and 319 reported an association with asthma hospitalization among children. Watanabe et al. (2016) 320 analyzed the same LIDAR data and reported reduced PEF in asthmatic adults attributable to 321 heavy AD days. These findings indicate the potential impact of AD on asthmatics.

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323 The current study contributes to existing evidence in several ways. First, it showed that 324 severe asthmatics with active treatment were susceptible to specific chemical 325 constituent/source of  $PM_{2.5}$ . Second, to our knowledge, the study is one of the first to 326 investigate the health impact of exposures on asthmatics in an environment with generally 327 low air pollutant concentrations. Lastly, the study reported possible effect modification by 328 AD which increased the risk of exposure to specific particle source.

329

330 There are several limitations worth noting. Our exposure data came from a fixed monitoring 331 station with no spatial information. The analysis was therefore unable to capture the spatial 332 heterogeneity of chemical constituents or source-apportioned concentrations. Single-station 333 measurements, which represent the average exposure levels in a population, were assigned as 334 personal exposure in the current study. This might result in Berkson-type measurement error 335 that produces little to no bias but reduces statistical power (Zeger et al., 2000). Another 336 limitation is the potential confounding by covarying constituents. Our current method of 337 adjustment did not account for potential confounding by covarying constituents, which could 338 be addressed using different methods, for example, model residuals although results would

339 not be directly comparable (Mostofsky et al., 2012). Our study did not have information on 340 indoor air pollution. Hence, for constituents with indoor sources, the discrepancy between the 341 true ambient concentration and average personal exposure may introduce a classical error that 342 can result in underestimation of associations (Mostofsky et al., 2012). We also had no 343 information on airborne pollen which was prevalent during the spring season and might 344 trigger/aggravate asthma symptoms. Further research is necessary to understand the 345 mechanism of how air pollen might influence exposure to PM<sub>2.5</sub> constituents/source profiles. 346 There was no data on daily mobility and activity of study participants, and no complete daily 347 information on their respiratory symptoms, medications, and personal characteristics (e.g. height and weight) which might be necessary to account for between-subject variability. The 348 349 source profiles resolved in the current study location might have unique particle mixtures that 350 limit interpretability in another location, despite the same labels. Some chemical constituents 351 might act as surrogates for other air pollutants including those not examined in this study. We 352 have tested the sensitivity of effect estimates in multi-pollutant models, but the adjustment 353 for other chemical species/sources might have resulted in a multicollinearity problem, 354 making it difficult to identify the significant chemical species/sources given their correlation 355 with the  $PM_{2.5}$  total mass.

356

#### 357 **5.** Conclusions

We conducted a study to examine the association of PM<sub>2.5</sub>, its constituents, and sources with 358 359 the lung function of severe asthmatic adults in urban Nagasaki, Japan. There was evidence of 360 lung function deterioration associated with sulfate particles and PM<sub>2.5</sub> from oil combustion and traffic source. OC and PM<sub>2.5</sub> from secondary sulfate source were also associated with 361 362 decreased respiratory capacity even though their estimates were not robust to the adjustment 363 of other pollutants. Findings suggest that despite their ongoing treatment and the overall low 364 air pollution level, asthmatic adults might still be affected by specific PM<sub>2.5</sub> constituents and 365 sources. Further research with larger sample size, longer study period, and more personal 366 exposure information to confirm the results will be beneficial.

367

#### 368 **Financial interest**

369 All authors have no competing financial interest.

370

#### 371 Acknowledgements

#### SCRIPT

372	The authors would like to thank Masahiko Araki, Atsuko Mori and Takuma Maeda								
373	(Nagasaki Prefectural Institute of Environment and Public Health, Nagasaki, Japan) for their								
374	assistance, and Seung-Muk Yi (Seoul National University, South Korea) for his advice. This								
375	study was supported by Nagasaki Prefectural Government, and partly supported by a research								
376	grant from the Non-profit Organization Aimed to Support Community Medicine Research in								
377	Nagasaki. CFS Ng was funded by a research fellowship from the Universities UK								
378	International (UUKi) Rutherford Fund Strategic Partner Grants.								
379									
380	Appendix A. Supplementary data								
381	Supplementary material is available at http://dx.doi.org/XXXXX.								
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#### Table 1 Characteristics of study subjects. 599

600

	Male (n = 11)				<b>Female</b> (n = 24)					
Characteristic <sup>a</sup>	Mean	SD	Median	Min	Max	Mean	SD	Median	Min	Max
Age (years)	52.7	10.8	47	38	65	50.4	12.8	51	20	65
Lung Function										
Number of measurement (day)	199.4	87.5	186	31	276	212.6	58.7	226.5	78	276
PEF (L/min)	499.3	86.4	530	280	690	328.7	88.2	340	130	640
Monthly maximum PEF (L/min)	516.1	76.8	530	380	690	355.9	91.8	360	160	640
Decrease from monthly maximum (%)	-4.3	4.7	-3.4	-46.2	0	-8	7.5	-6.1	-51.7	0
Decrease $\geq 15\%$ (event/person)	6.9	14.7	0	0	49	28.9	46.7	3.5	0	158
First decrease $^{b} \ge 15\%$ within a week										
period (event/person)	1.6	2.4	0	0	8	3.7	3.4	2	0	10
<sup>a</sup> All non-smokers. <sup>b</sup> include morning and evening measurements			R							

**Table 2** Summary of fine particulate matter (PM<sub>2.5</sub>), its chemical constituents, sources, and

- 602 weather variables.

	Mean	SD	Median	IQR	% of PM25 total mass
PM <sub>2.5</sub> (μg/m <sup>3</sup> )	18.5	8.34	17.79	10.13	606 60 <b>7</b>
Carbon (µg/m <sup>3</sup> )					608
OC	2.82	1.34	2.60	1.80	6992
EC	1.17	0.80	1.00	0.77	6403
<b>Ion</b> (μg/m <sup>3</sup> )					611
SO <sub>4</sub> <sup>2-</sup>	5.44	2.93	5.00	3.70	612 6134
$\mathrm{NH_4}^+$	2.36	1.41	2.20	1.70	61248
NO <sub>3</sub>	0.95	1.37	0.46	1.14	6 <u>3</u> 5 <u>1</u>
Metal (µg/m <sup>3</sup> )					616 617
Al	0.0711	0.0638	0.0530	0.0635	06384
As	0.0014	0.0012	0.0011	0.0010	060098
Cd	0.0002	0.0002	0.0002	0.0001	0.6001
Cu	0.0023	0.0016	0.0022	0.0016	$0.01^{1}_{.013}$
Fe	0.0899	0.0619	0.0820	0.0560	0486
K	0.0001	0.0001	0.0001	0.0001	062041
Mn	0.0056	0.0040	0.0047	0.0042	05230
Ni	0.0013	0.0008	0.0011	0.0007	0.9097
Pb	0.0078	0.0077	0.0057	0.0047	0,042
Ti	0.0063	0.0046	0.0053	0.0045	062394
V	0.0026	0.0019	0.0021	0.0016	0663104
Zn	0.0259	0.0207	0.0220	0.0190	0.940 632
Source (µg/m <sup>3</sup> )					633
Oil combustion & traffic	2.30	1.96	1.95	2.64	12.4 634
Dust & soil	3.42	2.96	2.72	3.05	18.5
Secondary sulfate	7.09	5.12	6.29	6.65	<b>63</b> 53
Secondary nitrate	2.21	3.57	0.82	2.86	6369
Sea salt	1.01	1.08	0.76	1.00	6 <i>3</i> 7
Industrial	1.82	2.57	1.05	1.66	9.8 638
Weather					050
Temperature (°C)	15.8	4.4	16.5	5.9	639
Relative humidity (%)	69.7	12.3	68	17	640-

#### 642 Figure Legends

- 643
- 644 **Fig. 1.** Percent change of PEF rate from personal monthly maximum (top panel) and odds
- ratios of the first 15% or more decrease of PEF from personal monthly maximum (bottom
- 646 panel) within a rolling 1-week interval for an IQR increase of PM<sub>2.5</sub> total mass or chemical
- 647 constituents at lag 0, lag 1, and the average denoted by "C" or cumulative lag. Refer to Table
- 648 2 for the IQR values.
- 649
- **Fig. 2.** Percent change of PEF rate from personal monthly maximum (top panel) and odds
- ratios of the first 15% or more decrease of PEF from personal monthly maximum (bottom
- 652 panel) for an IQR increase of the PMF-resolved source-specific PM<sub>2.5</sub> concentration at lag 0,
- lag 1, and the average denoted by "C" or cumulative lag. Refer to Table 2 for the IQR values.

CEP (E)





### Highlights:

- Lung function of asthmatic adults was inversely associated with sulfate in PM<sub>2.5</sub>.
- PM<sub>2.5</sub> from oil combustion & traffic reduced peak expiratory flow rate (PEF).
- Reduction associated with PM<sub>2.5</sub> from dust & soil was larger on Asian Dust days.
- No evidence linking total mass and selected metals to reduced lung function.
- Sensitivity to specific constituents/sources was suggested despite low PM<sub>2.5</sub> level.

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#### **Declaration of interests**

 $\boxtimes$  The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: